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Temporal dynamics of *Prochlorococcus* ecotypes in the Atlantic and Pacific oceans

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To better understand the temporal and spatial dynamics of Prochlorococcus populations, and how these populations co-vary with the physical environment, we followed monthly changes in the abundance of five ecotypes—two high-light adapted and three low-light adapted—over a 5-year period in coordination with the Bermuda Atlantic Time Series (BATS) and Hawaii Ocean Time-series (HOT) programs. Ecotype abundance displayed weak seasonal fluctuations at HOT and strong seasonal fluctuations at BATS. Furthermore, stable 'layered' depth distributions, where different Prochlorococcus ecotypes reached maximum abundance at different depths, were maintained consistently for 5 years at HOT. Layered distributions were also observed at BATS, although winter deep mixing events disrupted these patterns each year and produced large variations in ecotype abundance. Interestingly, the layered ecotype distributions were regularly reestablished each year after deep mixing subsided at BATS. In addition, Prochlorococcus ecotypes each responded differently to the strong seasonal changes in light, temperature and mixing at BATS, resulting in a reproducible annual succession of ecotype blooms. Patterns of ecotype abundance, in combination with physiological assays of cultured isolates, confirmed that the low-light adapted eNATL could be distinguished from other low-light adapted ecotypes based on its ability to withstand temporary exposure to high-intensity light, a characteristic stress of the surface mixed layer. Finally, total Prochlorococcus and Synechococcus dynamics were compared with similar time series data collected a decade earlier at each location. The two data sets were remarkably similar—testimony to the resilience of these complex dynamic systems on decadal time scales.

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Introduction

A key challenge facing marine microbiology is to understand how microbial diversity and biogeochemical cycles are linked, and to eventually incorporate this understanding into conceptual and predictive ocean models. Physiological and genetic analyses of cultured isolates, as well as metagenomic studies of whole communities (Venter et al., 2004; DeLong et al., 2006), are uncovering more and more about the metabolic potential of microbes comprising these assemblages. In addition, time series studies are revealing how the composition of microbial communities varies over both time and space (Fuhrman et al., 2006; Carlson et al., 2009; Treusch et al., 2009). Coupling the spatial and temporal dynamics of specific microbial groups with insights into their metabolic potential is essential for developing a quantitative understanding of the roles these microbes have in marine ecosystems.

Over the past 20 years, investigations of the unicellular cyanobacterium Prochlorococcus have provided insight into both the biogeography and metabolic potential of this group. Prochlorococcus is typically the most abundant photoautotroph in tropical and subtropical waters (Campbell et al., 1994; Partensky et al., 1999), and its abundance varies seasonally at some locations (Campbell et al., 1997; DuRand et al., 2001). Studies of cultured isolates have revealed the optimal light and temperature levels differ among the strains (Moore et al., 1998, 2002; Moore and Chisholm, 1999; Zinser et al., 2007), as do the nutrient pools available to them (Moore et al., 2002, 2005).

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Genomic analyses of isolates (Rocap et al., 2003; Coleman et al., 2006; Kettler et al., 2007), and metagenomic analyses of natural communities (DeLong et al., 2006; Martiny et al., 2006, 2009a), have provided additional insights into the genetic diversity, metabolic potential and evolutionary history of the group. The ability to examine the abundance and distribution of *Prochlorococcus* in the wild, and measure its genomic and metabolic properties in the lab and field, make Prochlorococcus a model system for advancing our understanding the ecology of marine microbes.

Prochlorococcus is composed of several clades that are physiologically and phylogenetically distinct with respect to their optimal light and temperature environments (Moore et al., 1998; Moore and Chisholm, 1999; Rocap et al., 2002; Johnson et al., 2006; Kettler et al., 2007; Zinser et al., 2007). These clades have been referred to as 'ecotypes' (Moore et al., 1998; Moore and Chisholm, 1999; Rocap et al., 2002) following the broader historical designation for genetically distinct subgroups within a species that are adapted to specific environments (Turesson, 1922; Clausen et al., 1940). They do not necessarily conform to the more recent and narrowly defined 'ecotype' concept developed by Cohan and others (Cohan, 2001; Cohan and Perry, 2007), which has become a notable model for exploring the theoretical basis for divergence among bacteria (Ward et al., 2006; Frasier et al., 2009). A more detailed discussion of the different uses of the term 'ecotype' is provided by Coleman and Chisholm, (2007).

The abundance of *Prochlorococcus* ecotypes in various oceanic regions has been studied extensively (West and Scanlan, 1999; Ahlgren et al., 2006; Bouman et al., 2006; Johnson et al., 2006; Zinser et al., 2006, 2007). Members of the two high-light adapted (HL) ecotypes, eMIT9312 and eMED4, are most abundant in the upper regions of the euphotic zone, whereas low-light adapted (LL) ecotypes such as eNATL and eMIT9313 are most abundant in the lower euphotic zone (West et al., 2001; Johnson et al., 2006; Zinser et al., 2007). These distribution patterns agree well with differences in the optimal light levels for representative ecotype strains (Moore and Chisholm, 1999; Zinser et al., 2007). Furthermore, eMED4 tends to dominate in cooler, higher latitude waters, whereas eMIT9312 dominates warmer, lower latitude waters (Johnson et al., 2006; Zwirglmaier et al., 2007); also in good agreement with temperature optima of representative strains. Water column stability and nutrient concentrations have also been correlated with ecotype abundance (Bouman et al., 2006; Johnson et al., 2006), and community structure (Martiny et al., 2009b), although the underlying causalities of these relationships are not well understood.

The environmental factors influencing *Prochloro*coccus ecotype abundance, such as light, temperature and water column mixing, vary over time and can display seasonal patterns, thus we might expect ecotype dynamics to do the same. Although we have analyzed ecotype variability over a span of a few days (Zinser et al., 2007), extensive time-series studies have not been conducted, and thus little is known about the dynamics of Prochlorococcus ecotypes on the scale of months to years. Analyses on these time scales should help refine our understanding of ecotype/environment interactions, and provide data sets for testing models designed to explore the dynamics of phytoplankton community structure (Follows et al., 2007).

To this end, we followed the spatial and temporal dynamics of *Prochlorococcus* ecotypes at monthly intervals over 5 years in coordination with the Bermuda Atlantic Time Series (BATS) and the Hawaii Ocean Time-series (HOT) programs (Karl and Lukas, 1996; Steinberg et al., 2001). Both programs are focused on oligotrophic, open ocean sites where *Prochlorococcus* is found in abundance (Campbell et al., 1994; DuRand et al., 2001). However, the physics and chemistry of these locations differ, most notably by the stronger seasonal mixing events at BATS and the higher inorganic phosphate concentrations at HOT (Wu et al., 2000; Cavender-Bares et al., 2001; Steinberg et al., 2001). Here, we explore how changes in environmental factors such as mixing, light and temperature are related to ecotype abundance and distribution, as well as how they co-vary temporally over 5 years. We also examine some of the emergent patterns in the field data through studies of light-shock tolerance in cultured isolates of different *Prochlorococcus* ecotypes.

Materials and methods

Sample collection

Beginning in November 2002, flow cytometry and qPCR samples were collected over a 5-year period during monthly cruises for the Bermuda Atlantic Time Series (BATS) and Hawaii Ocean Time-series (HOT) programs. Additional samples were collected bi-weekly between February and April at BATS. Samples were collected from 12 depths (1, 10, 20, 40, 60, 80, 100, 120, 140, 160, 180 and 200 m) at the BATS site (~ 5 nautical mile radius around 31° 40′N, 64° 10′ W), and from 12 depths (5, 25, 45, 60, 75, 85, 100, 115, 125, 150, 175 and 200 m) at Sta. ALOHA (~ 5 nautical mile radius around 22° 45'N, 158° 00'W). These locations are referred to BATS and HOT throughout the article for simplicity.

Flow cytometry

Whole seawater samples were immediately fixed with glutaraldehyde (final conc. 0.125% v/v) for 10 min, frozen in liquid nitrogen, and stored at -80 °C until samples could be processed in the laboratory using an influx flow cytometer



(Becton Dickinson, Franklin Lakes, NJ, USA). *Prochlorococcus* and *Synechococcus* populations were identified and quantified based on their unique autofluorescence and scatter signals (Olson *et al.*, 1990a, 1990b). *Prochlorococcus* could not always be clearly distinguished in the upper 40 m at BATS from June–September, thus these profiles were excluded from depth-integrated counts in Figure 2. Flow cytometry profiles from early 2003 were not processed.

Primer re-design for ecotype eSS120

New primers for the eSS120 ecotype were designed in ARB (Ludwig et al., 2004), using a large database of environmental ITS sequences (Martiny et al., 2009b). These primers, 5'-AACAAACTTTCTCCTG GGCT-3' and 5'-AGTTGATCAGTGGAGGTAAG-3', matched 87% of known eSS120 ITS sequences, but did not target MIT9211. The specificity of the primers was initially tested against cultured isolates from other ecotypes such as MIT9313, MIT9312, MED4 and NATL2a. These tests confirmed specificity within the dynamic range of the assay (~ 5 to 5×10^5 cells ml⁻¹). Specificity was also confirmed by cloning and sequencing ITS regions amplified from field samples from BATS and HOT using the new primers. All amplified ITS sequences clustered with strains SS120 and MIT9211, both members of the eSS120 ecotype (Kettler et al., 2007), in a bootstrapped (n = 100) neighbor-joining tree constructed with the Bosque software package (Ramirez-Flandes and Ulloa, 2008) (Supplementary Figure 1). The new primers substantially increased counts of the SS120 ecotype (Supplementary Materials).

Quantitative PCR

Samples for qPCR were collected and processed as described previously (Ahlgren et al., 2006; Zinser et al., 2006, 2007) with two small modifications. First, reaction volumes were reduced from 25 µl to 15 μl and performed in 384-well plates on the Light Cycler 480 for samples collected after 2003. Second, concentrations of eMIT9313-specific primers were increased to 5 µM to improve sensitivity. Estimated abundances that fell below the lowest value of the standard curve were set to the theoretical detection limit of 0.65 cells ml⁻¹. Samples were excluded if their melt curves contained multiple peaks or peaks different from those in the DNA standards to ensure only the targeted ecotypes were quantified. Missing data were determined by linear interpolation when abundance estimates were available for the depths immediately above and below the missing value.

Environmental data extraction

Temperature, salinity and potential density were downloaded from the HOT and BATS websites. Missing data were determined by linear interpolation when values were available for the depths immediately above and below the missing data point. Mixed layer depths, light attenuation coefficients and Prochlorococcus and Synechococcus abundance (1991–1995) were also downloaded directly from the HOT website. Mixed layer depths and attenuation coefficients at BATS were calculated from bottle-derived profiles and SeaWiFS Profiling Multichannel Radiometer profiles of photosynthetically active radiation (PAR) collected by Bermuda Bio-Optics Program. The mixed layer depth was determined when potential density differed from surface values by $> 0.125 \text{ kg m}^{-3}$. Attenuation coefficients at BATS were calculated by linear regression of logtransformed PAR values; only profiles with an $R^2 > 0.993$ were used.

Solar irradiance

Surface irradiance was determined from SeaWiFSderived estimates of daily-integrated PAR. Eight-day means of daily integrated PAR values from March 2000 to July 2006 were calculated from a 27 km by 27 km region around BATS and HOT (White et al., 2007). A two component Fourier model with a period of 1 year was fitted to PAR data, producing an R² of 0.90 and 0.93 at BATS and HOT, respectively. Modeled PAR data were used to estimate surface irradiance on the day of sample collection. This model was necessary to account for seasonal variability in solar flux due to changes in day length and solar azimuth. These changes result in a roughly two-fold difference in daily-integrated solar flux between summer and winter (for example, $\sim 20\,000-56\,000\,\mathrm{mE}\,\mathrm{m}^{-2}\,\mathrm{d}^{-1}$ at BATS, and $\sim 31\,000-58\,000\,\mathrm{mE}\,\mathrm{m}^{-2}\,\mathrm{d}^{-1}$ at HOT, using this model).

The relationships between ecotype abundance and PAR plotted in Figure 1 and Supplementary Figure 2 were determined using robust locally weighted linear regression (LOWESS) in MATLAB. Robust LOWESS is more resistant to outliers (Cleveland, 1979), which are defined in MATLAB's robust LOWESS function as data outside six mean absolute deviations.

Time series and other statistical analyses

Integrated ecotype abundance, surface PAR, and mixed layer depth were log-transformed, detrended, and resampled at a regular monthly interval to meet the mathematical requirements for time series analyses (Legendre and Legendre, 1998). Detrended data were calculated as the residuals of linear regression of log-transformed data against time. Detrended data were resampled every 30.44 days, which is equivalent to 12 measurements per year, using linear interpolation. Coefficients of autocorrelation and cross-correlation were determined in MATLAB, and s.d. calculated as $n^{-1/2}$, where n is the number or resampled data points. Spectral



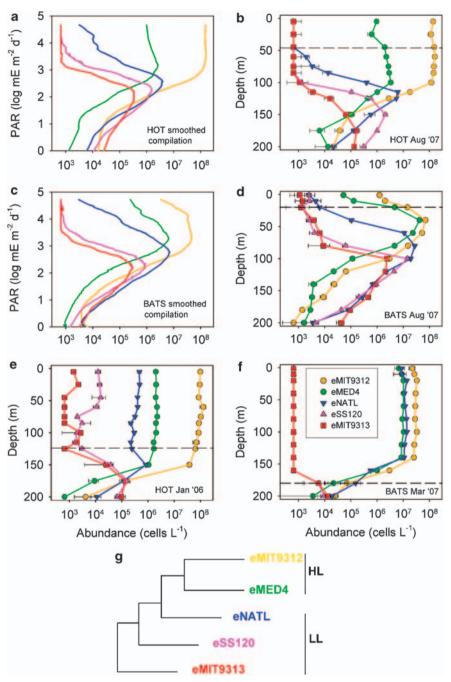


Figure 1 Ecotype distribution patterns in relation to depth and photosynthetically active radiation (PAR). A smoothed illustration of the relationship between ecotype abundance and irradiance, calculated from a compilation of data from all 5 years, was plotted for HOT (a) and BATS (c). Data were smoothed by locally weighted regression of ecotype abundance against irradiance. Smoothed data exclude profiles with a mixed layer depth > 100 m. Representative depth profiles of ecotype abundance at HOT (b, e) and BATS (d, f) during periods of stratification and deep mixing. Dashed lines mark the mixed layer depth, and error bars represent one s.d. An illustration of the color scheme and general phylogenetic relationships among ecotypes after (Rocap et al., 2002; Kettler et al., 2007) is displayed in (g). This tree serves only as reference to the other panels; branch lengths are not to scale.

analysis by discrete Fourier transformation calculated using the Fast Fourier Transformation (FFT) algorithm, was also performed in MATLAB. The power spectral density was estimated as the absolute value of FFT².

Differences among ecotypes in the average depth of maximum abundance were tested using repeatedmeasures ANOVA, followed by a Tukey post-test ($\alpha = 0.05$), using MATLAB.

Non-parametric partial correlation coefficients (Spearman R) were calculated in MATLAB to determine the relationship between abundance and temperature while controlling for the influence of light.

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Light shock experiments

Serial batch cultures of axenic *Prochlorococcus* strains MED4, NATL2a and SS120 were grown in Sargasso seawater-based Pro99 media (Moore *et al.*, 2007) and illuminated by cool white fluorescent lamps. Cultures were transferred at least four times to acclimatize them to $35\,\mu\text{E}\,\text{m}^{-2}\,\text{s}^{-1}$ of continuous light. Duplicate acclimated cultures that were in logphase growth were then exposed to $400\,\mu\text{E}\,\text{m}^{-2}\,\text{s}^{-1}$ for 4h before being returned to their initial light levels. *In vivo* chlorophyll fluorescence was measured before and after light shock with a 10-AU fluorometer (340–500 nm excitation and 680 nm emission filters), and cell counts determined by flow cytometry as described above.

Results and Discussion

Ecotype distribution with light and temperature Striking similarities between HOT and BATS emerged when data from all depths and all 5 years were combined for a synoptic analysis of *Prochloro*coccus ecotype abundance along irradiance/depth gradients (Figures 1a and b; Supplementary Figure 2). As expected, the two high-light adapted ecotypes, eMIT9312 and eMED4, were most abundant at higher irradiances, with abundance dropping off sharply below 1500 mE m⁻² d⁻¹ of photosynthetically active radiation (Figures 1a and b; Supplementary Figures 2a and e). In contrast, two low-light adapted ecotypes, eSS120 and eMIT9313, usually reached maximum abundances between 100 and $250\,mE\,m^{-2}\,d^{-1},$ and were typically at or near detection limits at irradiances $>\!1500\,mE\,m^{-2}\,d^{-1}.$ The eNATL group, also low-light adapted had an intermediate distribution, reaching maximum abundance at 300–600 mE m⁻² d⁻¹. Unlike the other low light ecotypes, eNATL abundance was occasionally high at irradiance levels $> 1500 \,\mathrm{mE}\,\mathrm{m}^{-2}\,\mathrm{d}^{-1}$ (Figures 1a and b; Supplementary Figures 2c and h), which is consistent with hypothesis that eNATL can tolerate exposure to higher light levels than eSS120 and eMIT9313 (Coleman and Chisholm, 2007; Zinser et al., 2007).

This consistent relationship between irradiance and ecotype abundance results in a 'layered' depth distribution at both locations (Figures 1a-d). At HOT, for example, HL ecotype eMIT9312 tended to reach maximum abundance at shallower depths than did its fellow HL ecotype eMED4 (Table 1). LL ecotypes also partitioned the water column at HOT, with eNATL abundance peaking at significantly shallower depths than eSS120 and eMIT9313 (Table 1). Similar patterns in ecotype distribution were also observed at BATS (Table 1; Figures 1b and d), except for during deep mixing events, defined here as when the mixed layer depth was >100 m. During periods of deep mixing physical homogenization appears to overwhelm biological partitioning of the water column, resulting in

Table 1 Average depth of maximum ecotype abundance at HOT and BATS (mean ± STD)

Ecotype	HOT (m)	BATS (m)
eMIT9312 eMED4 eNATL eSS120 eMIT9313	42 ± 27^{A} 71 ± 29^{B} 105 ± 18^{C} 118 ± 26^{D} 128 ± 30^{D}	54 ± 28^{A} 60 ± 26^{A} 87 ± 22^{B} 101 ± 28^{C} 107 ± 25^{C}

Samples collected when the mixed layer depth was $> 100\,\mathrm{m}$ were excluded. Values with different superscripts are significantly different (Tukey post test of repeated measures ANOVA; $\alpha=0.05$).

uniform depth distributions (Figures 1e and f). Therefore, data from periods of deep mixing were removed from statistical analysis of depth distributions.

It is remarkable that, with the exception of periods of deep mixing, the general patterns of ecotype abundance appear relatively consistent in the Atlantic and Pacific despite substantial differences in the chemical and physical environment. Furthermore, these patterns are consistent with those found previously in a variety of ocean regions (West and Scanlan, 1999; West et al., 2001; Bouman et al., 2006; Johnson et al., 2006; Zinser et al., 2006, 2007). That is, the ecotypes tend to partition the water column by depth, with eNATL reaching maximum abundance in between the peaks of HL ecotypes MED4 and MIT9312, and other LL ecotypes. The similarities in the distributions along depth/light gradients suggest that Prochlorococcus ecotypes are responding in a consistent fashion to irradiance regardless of geography.

Temporal dynamics: depth-integrated Prochlorococcus and Synechococcus populations

Using flow cytometry, we measured the abundance of Prochlorococcus, and its close relative Synechococcus (Rocap et al., 2002), to provide an overall framework for exploring *Prochlorococcus* ecotype dynamics. At BATS, the depth-integrated *Prochloro*coccus population displayed a strong seasonal pattern, reaching the highest levels in the late summer and fall, and the lowest in the late winter during the annual deep mixing events (Figure 2a). Synechococcus displayed the inverse pattern, and even occasionally exceeded the abundance of *Prochlorococcus* during deep mixing events (Figure 2b). This is the same pattern reported by DuRand et al, (2001) for Prochlorococcus and Synechococcus at BATS from 1990 to 1994 (Figures 2a and b). The concordance between these two data sets is remarkable, given that they are separated by more than a decade.

Variations in *Prochlorococcus* and *Synechococcus* abundance were much less dramatic at HOT. Integrated abundance of *Prochlorococcus* varied just over 2-fold throughout the time series, and did not always reach maximal abundance in the summer or minimal abundance in the winter (Figure 2c).



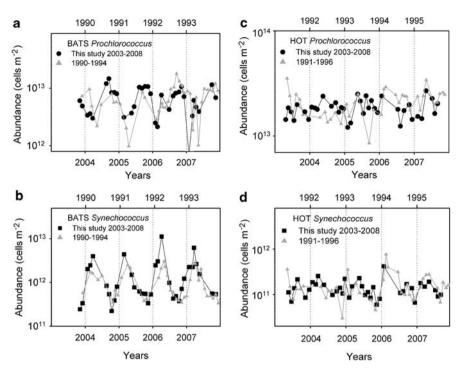


Figure 2 Abundance of Prochlorococcus and Synechococcus at BATS and HOT determined by flow cytometry (bottom x axis). Abundance levels from this study are compared with those collected at BATS from 1990 to 1994 (DuRand et al., 2001) (a, b), and at HOT from 1991 to 1995 (c, d) (top x axis), which were also determined by flow cytometry.

Synechococcus still displayed an annual pattern, tending to reach peak abundance during winter, but they were never as abundant as Prochlorococcus, in contrast to what was observed at BATS. These abundance and variability levels are also consistent with those observed over a decade ago at HOT (Figures 2c and d)

Temporal dynamics: depth-integrated ecotype abundance patterns

As was seen in the total Prochlorococcus population, the depth-integrated (0-200 m) abundance of all five ecotypes followed clear annual patterns at BATS (Figure 3a). Spectral analysis of each ecotype revealed dominant peaks in the power spectrum at a period of 1 year (Supplementary Figure 3), and autocorrelations displayed a sinusoidal pattern, with peaks in autocorrelation every 12 months (Supplementary Figure 4). Furthermore, fitting a single component Fourier series with a period of 1 year to each ecotype produced R^2 values ranging from 0.48 (eMIT9313) to 0.67 (eSS120), indicating that most of the variability in integrated abundance could be accounted for by an annual oscillation.

Annual variations in temperature and mixed layer depth were smaller at HOT than at BATS, as were the variations in ecotype abundance (Figure 3b). Although spectral analysis did reveal that all ecotypes had a peak in the power spectrum at a period of 1 year, this was not the only strong signal,

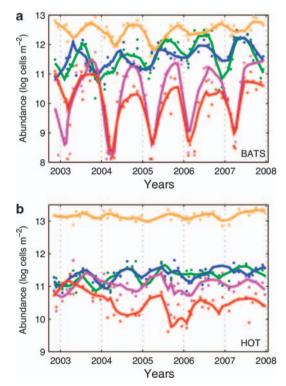


Figure 3 Integrated ecotype abundance (0-200 m) at BATS (a) and HOT (b) from 2003 to 2008. Solid lines represent abundances smoothed by locally weighted regression for eMIT9312 (yellow), eMED4 (green), eNATL (blue), eSS120 (purple) and eMIT9313 (red).



particularly for the LL ecotypes (Supplementary Figure 5). Fitting a single component Fourier series with a period of 1 year to each ecotype produced R^2 values ranging from only 0.17 (eMIT9312) to 0.33 (eMIT9313). In addition, autocorrelations did not display the same strong sinusoidal pattern as seen at BATS (compare Supplementary Figure 3 with Supplementary Figure 6). Thus, while there was a component of annual variability to ecotype abundance at HOT, there was also a variability at periods greater and less than 1 year. This suggests that when annual environmental variations are more moderate, the impact of intra- and inter-annual events, such as passage of mesoscale eddies or El Niño/La Niña oscillations—both known to influence primary production and phytoplankton community composition at HOT (Karl et al., 1995; Letelier et al., 2000; Corno et al., 2007; Bibby et al., 2008)—may become more pronounced.

Although ecotype abundance followed a clear annual cycle at BATS, the cycles were not synchronized among ecotypes, that is, different ecotypes reached peak abundance at different times. In each of the 5 years, the eNATL ecotype reached its maximum integrated abundance about 4 months after winter deep mixing events (Table 2), typically in June (Supplementary Figure 7). Abundance of eMED4 peaked roughly 1 month later, whereas the LL-clades eSS120 and eMIT9313 reached maximal abundance about 8 months after the deep mixing event (Table 2). HL ecotype eMIT9312 also reached maximal abundances at around the same time as eSS120 and eMIT9313, typically in October or November (Supplementary Figure 7). The regularity of this pattern indicates that each ecotype responded to changes in environmental conditions in different, yet consistent ways, resulting in an annually repeating succession of ecotypes.

The succession of *Prochlorococcus* ecotypes is in some ways reminiscent of classical phytoplankton succession models, although it is occurring at a much smaller phylogenetic scale; all *Prochlorococcus* differ in 16S rRNA sequence by <3% (Moore *et al.*, 1998), which would collectively constitute a single

Table 2 Correlation and cross-correlation between integrated ecotype abundance (0–200 m) and mixed layer depth at BATS

Ecotype	Correlation coeff. (lag in months)	Max. cross-correlation coeff. (lag in months)
eMIT9312	-0.06 (0)	0.61* (8)
eMED4	-0.62* (0)	0.58* (5)
eNATL	-0.54* (0)	0.65* (4)
eSS120	-0.60* (0)	0.66* (8)
eMIT9313	-0.46* (0)	0.61* (8)

The coefficient of cross-correlation represents the relationship between abundance at 1 month and the mixed layer depth from previous months. The maximum positive cross-correlation is reported alongwith the delay in months (lag) between abundance and mixed layer depth when correlation is greatest. *Indicates P < 0.05.

bacterial species by conventional standards (Stackebrandt and Goebel, 1994). Interestingly, two recent time series studies at BATS have also uncovered annual cycles and succession patterns in other microbial groups, most notably the SAR11 clade (Carlson et al., 2009; Treusch et al., 2009). For example, one SAR11 subgroup reaches peak abundance in surface waters during the summer, whereas another subgroup peaks in the winter (Carlson et al., 2009). SAR11 bacteria are the most abundant heterotrophs at BATS and are major consumers of dissolve organic compounds (Morris et al., 2002; Malmstrom et al., 2005), whereas Prochlorococcus is the most abundant photoautotroph and a substantial source of dissolved organics (DuRand et al., 2001; Bertilsson et al., 2005). In addition, SAR11 bacteria and Prochlorococcus are both major consumers of small compounds like amino acids and dimethylsulfoniopropionate (DMSP) (Zubkov et al., 2003; Malmstrom et al., 2004; Vila-Costa et al., 2006; Michelou et al., 2007), which are significant sources of C, N and S to marine microbial communities. Therefore, it seems plausible that succession in these two abundant groups could be linked through the production of, and competition for, dissolved organic compounds. Uncovering potential links in the dynamics of the dominant microbial groups presents a future challenge.

Ecotype abundance in different regions of the euphotic zone

Variations in light and temperature levels, which impact the growth of *Prochlorococcus* (Moore and Chisholm, 1999; Johnson *et al.*, 2006; Zinser *et al.*, 2007), are greater in surface waters than at depth. Thus integrating abundance across the entire water column likely obscures important features in ecotype dynamics. To get a more detailed understanding of the temporal and spatial dynamics of the ecotypes, we analyzed integrated abundance in three sections of the euphotic zone (0–60 m, 60–120 m and 120–200 m).

The HL-adapted ecotype eMIT9312 displayed similar abundance patterns in the upper $(0-60\,\mathrm{m})$ and middle $(60-120\,\mathrm{m})$ euphotic zone at BATS, but below 120 m the seasonal cycle was out of phase with surface cycles by several months (Figure 4a; Figure 5a). Abundance in the lower euphotic was positively correlated with mixed layer depth (Spearman R=0.5; P<0.05), and abundance peaks occurred simultaneously with annual deep mixing events (Figure 4a). This suggests that it is the transport of surface populations below 120 m, and not in situ growth that may be responsible for most of the annual variations in abundance of eMIT9312 in the lower euphotic zone at BATS.

As with eMIT9312, patterns of eMED4 abundance in the upper and middle euphotic zone also differed from those in the lower zone at BATS (Figure 5b), with strong spikes in abundance below 120 m

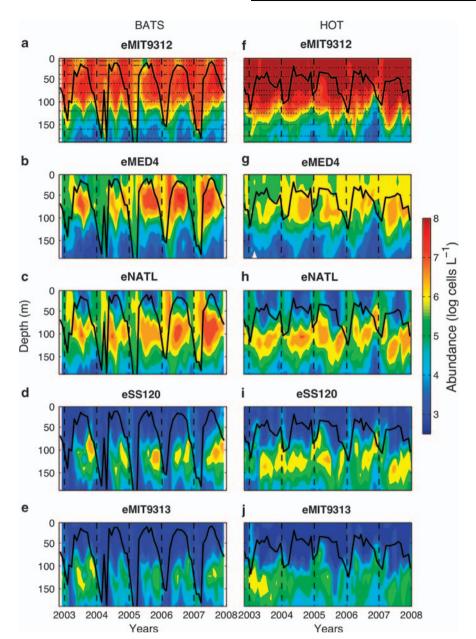


Figure 4 Ecotype abundance at BATS (a-e) and HOT (f-j) from 2003 to 2008. Sampling depths at HOT and BATS are indicated by the points overlaying (a and f). The solid lines indicate the mixed layer depth.

accompanying deep mixing (Figure 4b). But while the eMED4 and eMIT9312 dynamics were synchronized below 120 m, these HL-adapted ecotypes were out of synchronization by several months in the upper 120 m (Figures 5a and b). That is, each year at BATS, eMED4 typically reached peak abundance in July-August, whereas eMIT9312 reached maximum abundance in October-November (Supplementary Figure 7). At HOT, in contrast, these two HL ecotypes did not display these offset repeating patterns.

We hypothesize that different temperature sensitivities of eMED4 and eMIT9312 explain, at least in part, the differences in their temporal dynamics at BATS. That is, strains belonging to the eMED4 clade have lower temperature optima than those belonging to the eMIT9312 clade (Johnson et al., 2006; Zinser et al., 2007). If this differential trait is universal among cells belonging to the eMED4 and eMIT9313 clades, then this would allow eMED4 cells to accumulate during the first half of the year when temperatures were low, whereas the higher temperature optimum of eMIT9312 would limit their accumulation until later in the season when temperatures were high—as was observed at BATS. In fact, the abundance of eMED4 in the upper 60 m was negatively correlated with temperature when light levels were taken into account (partial correlation coeff. -0.43; P < 0.05), whereas eMIT9312 abundance was positively correlated with temperature

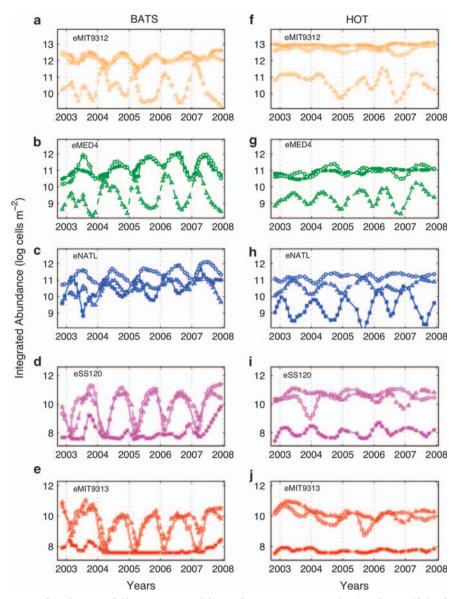


Figure 5 Integrated ecotype abundance in different regions of the euphotic zone at BATS (a-e) and HOT (f-j). The regions from 0-60 m, 60-120 m and 120 m are identified by filled squares, open circles and open triangles, respectively. Abundance was smoothed using locally weighted regression to eliminate low frequency variation.

(partial correlation coeff. 0.30; P < 0.05). The potential influence of temperature on the temporal distribution of HL ecotypes at BATS is analogous to its inferred influence on their geographic distribution: eMED4 dominates in cooler, higher latitude waters, and eMIT9312 dominates warmer, lower latitude waters (Johnson et al., 2006; Zwirglmaier et al., 2007).

Abundance patterns of eNATL also differed among regions of the euphotic zone, and the similarity of these patterns between BATS and HOT was striking. At both locations, abundance in the upper 60 m typically peaked 3–4 months earlier than at 60–120 m and 120–200 m (Figures 4c and h; Figures 5c and h). Integrated abundance in the upper 60 m reached maximum levels during deep

mixing events, and decreased as the mixed layer depth shoaled (Figures 4c and h). This suggests that annual peaks in abundance in the upper 60 m were due, at least in part, to vertical transport of deeper cells to surface waters via mixing. While it remains unclear if members of the LL-adapted eNATL clade were able to grow at high light levels found in the upper euphotic zone, the net accumulation eNATL cells throughout the water column during periods of deep mixing confirmed that the eNATL clade was able to at least tolerate temporary exposure to high irradiance.

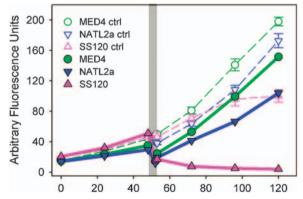
In contrast to eNATL, the abundance of the other two LL-adapted ecotypes, eSS120 and eMIT9313, did not increase dramatically in the upper 60 m during periods of deep mixing. In fact, their

abundance was typically near or below detection levels throughout the entire water column at BATS when the well-mixed layer spanned the euphotic zone (Figures 4d and e). It appears that, unlike eNATL, members of the eSS120 and eMIT9313 ecotypes could not tolerate temporary high light exposure when transported to surface waters via mixing. At HOT, however, the mixed layer depth rarely exceeded 100 m, and therefore a substantial fraction of eSS120 and eMIT9313 cells were not transported to surface waters during winter mixing events. Thus, it appears that eSS120 and eMIT9313 maintained relatively stable abundance levels throughout the time series, in comparison with BATS, as they were not subjected to inhibitory, or lethal, high light exposure during mixing events.

Response to light shock in cultured isolates

To directly test the hypothesis that members of the eNATL clade can withstand temporary exposure to high intensity light better than members of the other LL-ecotypes, we performed light shock experiments with cultured isolates from the eMED4, eNATL and eSS120 clades. Strains acclimated to an irradiance of $35 \,\mu\text{E}\,\text{m}^{-2}\,\text{s}^{-1}$ were exposed to light levels of 400 μE m⁻² s⁻¹ for 4 h and then returned back to $35 \,\mu\text{E}\,\text{m}^{-2}\,\text{s}^{-1}$. For perspective, these light levels are equivalent to those that would be found at 90 m and 36 m at mid-day, assuming a surface irradiance of $2000 \,\mu\text{E}\,\text{m}^{-2}\,\text{s}^{-1}$ and an attenuation coefficient of – 0.045 m⁻¹, which are typical of HOT and BATS. Furthermore, $400 \,\mu\text{E}\,\text{m}^{-2}\,\text{s}^{-1}$ is a lethal intensity for NATL2a and SS120, but not MED4, when applied continuously (Moore and Chisholm, 1999).

In vivo chlorophyll fluorescence of all three cultures dropped significantly relative to the control after the light shock. Fluorescence in MED4 and NATL2a cultures rebounded within 24 h and continued to increase over the next 48 h (Figure 6), whereas fluorescence of the SS120 culture contin-



6 Response of *Prochlorococcus* strains NATL2a and SS120 to light-shock. In vivo chlorophyll fluorescence of duplicate cultures acclimated to $35\,\mu\mathrm{E}\,\mathrm{m}^{-2}\,\mathrm{s}^{-1}$, exposed to 400 µE m⁻² s⁻¹ for 4 hours (gray area), and returned to $35\,\mu\text{E}\,\text{m}^{-2}\,\text{s}^{-1}.$ Controls did not experience light shock. Error bars represent one s.d.

ued to decline for the following 72 h. Cell concentrations of MED4 and NATL2a also rose following the light shock, whereas the number of SS120 cells declined throughout the recovery period (Supplementary Figure 8). These results indicate that NATL2a can withstand temporary light shock better than SS120, which is consistent with the temporal and spatial distribution patterns of their respective ecotypes at HOT and BATS described above. These results are also consistent with a similar light shock experiment comparing isolates SS120 and PCC 9511, a member of the HL-adapted eMED4 ecotype. In this study, the capacity to repair damage to photosystem II caused by light shock was much greater in PCC 9511 than in SS120 (Six et al., 2009). Although it was not tested specifically, it is plausible that NATL2a is also better able to reverse photoinactivation of photosystem II than SS120. If so, then this would present a possible mechanism to explain, at least in part, the differences in temporary light shock tolerance and the subsequent environmental distributions of LL ecotypes.

Analysis of the genomes of cultured isolates provides some insight into the genes that may be responsible for differences in light physiology among LL-adapted ecotypes. For example, the two sequenced eNATL isolates each have 41 genes encoding high light-inducible proteins (HLIP), whereas the other LL-adapted ecotypes only have 9-13 HLIP genes (Coleman and Chisholm, 2007). This protein family, also called small cab-like proteins, aids in high-light survival and photoacclimation in Synechocystis (He et al., 2001; Havaux et al., 2003), and may also have a role in light-shock tolerance in eNATL. The mechanism by which they do this remains unclear, but high light-inducible proteins are thought to physically associate with either photosystem and allow it to shed excess energy as heat and thereby reduce photoinactivation (Promnares et al., 2006; Yao et al., 2007). As a photosynthetic organism, reducing and reversing the effects of photosystem inactivation is crucial for the survival of *Prochlorococcus* exposed to high light levels.

Differences in their ability to repair UV-damaged DNA may also partially explain differences in the environmental distributions of eNATL and other LL ecotypes. The six sequenced isolates from the two HL-adapted ecotypes and the two sequenced isolates from the eNATL ecotype contain genes that encode photolyase (Coleman and Chisholm, 2007; Kettler et al., 2007), an enzyme helps repair UV-damaged DNA (Sancar, 2000). Photolyase genes are absent in the genomes of the other four LL-adapted isolates that have been sequenced. Although these other LL-adapted isolates do encode for an alternate putative UV-repair enzyme, pyrimidine dimer glycosylase (Goosen and Moolenaar, 2008; Partensky and Garczarek, 2010), its function in *Prochlorococcus* has not been confirmed. This enzyme also has a modification of the Arg-26



residue that is known to dramatically reduce activity in homologs (Doi et al., 1992; Goosen and Moolenaar, 2008), suggesting a diminished ability to repair UV-damaged DNA in eSS120 and eMIT9313 ecotypes. Indeed, the high-light adapted strain MED4, which encodes photolyase, has a greater tolerance to UV exposure than low-light adapted strain MIT9313 (Osburne et al., 2010), which lacks photolyase but encodes pyrimidine dimmer glycosylase. Thus, the protection from UV exposure provided by photolyase may explain, at least in part, why the eNATL clade can better survive transport to UV-rich surface waters.

Conclusions

Clear patterns in the temporal and spatial distribution of *Prochlorococcus* ecotypes emerge from this study. For example, ecotype abundance follows a strong annual pattern at BATS, whereas ecotype abundance has only a weak annual pattern at HOT. In addition, ecotypes at BATS follow an annual succession pattern, but a similar pattern is not observed at HOT. These patterns are consistent with what we have learned from physiological assays on cultured isolates with regards to the light optima, temperature optima, and light-shock tolerance of different ecotypes. That is, the distinct distribution patterns at both HOT and BATS can be explained, at least in general terms, by the consistent and predictable responses of ecotypes to changes in the light, temperature and mixing at each location.

Analyses of inorganic nutrient concentrations and ecotype abundance were not possible as nutrient levels were below detection in the upper 100 m at BATS throughout most of this study. However, while nutrients undoubtedly influence growth rates and standing stocks of *Prochlorococcus*, their influence on specific ecotypes may not be apparent, as evidenced by the fact that general patterns can be explained without them. Indeed, recent work suggests that nitrate concentrations might impact the composition of *Prochlorococcus* populations at finer phylogenetic levels than the ecotypes examined in our study (Martiny et al., 2009b). Exploring how, and at which temporal, spatial and phylogenetic scales, the chemical and biological environment influence Prochlorococcus abundance and diversity presents a future challenge.

Results from this study help set the stage for coupling patterns in temporal and spatial dynamics of ecotypes with insights into their metabolic potential. The next step is to understand how *Prochlorococcus* ecotypes, or even sub-groups within these ecotypes, might differ in terms of nutrient usage, dissolved organic matter production and consumption, and other metabolic processes. This understanding will come from additional studies involving strain isolation, metagenomic comparisons, and large-scale single-cell genomics.

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